The persistence of carbon in the African forest understory

Citation for published version:

Digital Object Identifier (DOI):
10.1038/s41477-018-0316-5

Link:
Link to publication record in Edinburgh Research Explorer

Document Version:
Peer reviewed version

Published In:
Nature Plants

General rights
Copyright for the publications made accessible via the Edinburgh Research Explorer is retained by the author(s) and / or other copyright owners and it is a condition of accessing these publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy
The University of Edinburgh has made every reasonable effort to ensure that Edinburgh Research Explorer content complies with UK legislation. If you believe that the public display of this file breaches copyright please contact openaccess@ed.ac.uk providing details, and we will remove access to the work immediately and investigate your claim.
Title

The persistence of carbon in the African forest understory

Authors

Wannes Hubau 1,2,3 *, Tom De Mil 1,2 *, Jan Van den Bulcke 2,4, Oliver L. Phillips 3, Bhély Angoboy Ilondea 1,5,6, Joris Van Acker 2,4, Martin J. P. Sullivan 3, Laurent Nsenga 1, Benjamin Toirambe 1, Camille Couralet 1, Lindsay F. Banin 7, Serge K. Begne 8,3, Timothy R. Baker 3, Nils Bourland 1,9,10,11, Eric Chezeaux 12, Connie J. Clark 13, Murray Collins 14, James A. Comiskey 15,16, Aida Cuni-Sanchez 17,18, Victor Deklerck 2,4, Sofie Dierickx 1, Jean-Louis Doucet 10, Corneille E. N. Ewango 19,20,21, Ted R. Feldpausch 22, Martin Gilpin 3, Christelle Gonmadje 23, Jefferson S. Hall 24, David J. Harris 25, Olivier J. Hardy 26, Marie-Noel D. Kamdem 8,27, Emmanuel Kasongo Yakusu 1,21,2, Gabriela Lopez-Gonzalez 7, Jean-Remy Makana 19, Yadvinder Malhi 28, Faustin M. Mbayu 21, Sam Moore 28, Jacques Mukanzi 19,29, Georgia Pickavance 3, John R. Poulsen 13, Jan Reitsma 30, Mélissa Rousseau 1,11, Bonaventure Sonké 8, Terry Sunderland 9,31, Hermann Taedoumg 8, Joey Talbot 3, John Tshibamba Mukendi 1,21,32, Peter M. Umunay 33, Jason Vleminckx 26,34, Lee J. T. White 35,36,37, Lise Zemagho 8, Simon L. Lewis 3,17, Hans Beeckman 1

* These authors contributed equally to this work.

Contact information

whubau@gmail.com, wannes.hubau@africamuseum.be, tom.demil@ugent.be

Service of Wood Biology, Royal Museum for Central Africa
Tervuren, Belgium, Leuvensesteenweg 13, 3080 Tervuren

Affiliations

1 Service of Wood Biology, Royal Museum for Central Africa, Tervuren, Belgium

2 UGent-Woodlab, Laboratory of Wood Technology, Department of Environment, Ghent University,
Ghent, Belgium
<table>
<thead>
<tr>
<th></th>
<th>Institution</th>
</tr>
</thead>
<tbody>
<tr>
<td>28</td>
<td>School of Geography, University of Leeds, Leeds, UK</td>
</tr>
<tr>
<td>29</td>
<td>Centre for X-ray Tomography (UGCT), Ghent University, Ghent, Belgium</td>
</tr>
<tr>
<td>30</td>
<td>Institut National pour l'Etude et la Recherche Agronomique, Kinshasa I, Democratic Republic of Congo</td>
</tr>
<tr>
<td>31</td>
<td>Ecole Régionale Postuniversitaire d'Aménagement et de Gestion intégrés des Forêts et Territoires tropicaux (ERAIFT), Kinshasa, Democratic Republic of Congo</td>
</tr>
<tr>
<td>34</td>
<td>Centre for Ecology and Hydrology, Penicuik, UK</td>
</tr>
<tr>
<td>35</td>
<td>Plant Systematic and Ecology Laboratory, Higher Teachers’ Training College, University of Yaounde I, Cameroon</td>
</tr>
<tr>
<td>36</td>
<td>CIFOR, Bogor, Indonesia</td>
</tr>
<tr>
<td>37</td>
<td>Forest Resources Management, Gembloux Agro-Bio Tech, University of Liege, Belgium</td>
</tr>
<tr>
<td>38</td>
<td>Resources and Synergies Development, Singapore, Singapore</td>
</tr>
<tr>
<td>39</td>
<td>Rougier-Gabon, Libreville, Gabon</td>
</tr>
<tr>
<td>40</td>
<td>Nicholas School of the Environment, Duke University, Durham, NC, USA</td>
</tr>
<tr>
<td>41</td>
<td>Grantham Research Institute on Climate Change and the Environment, London, UK</td>
</tr>
<tr>
<td>42</td>
<td>Inventory &amp; Monitoring Program, National Park Service, Fredericksburg, VA, USA</td>
</tr>
<tr>
<td>43</td>
<td>Smithsonian Institution, Washington, DC, USA</td>
</tr>
<tr>
<td>44</td>
<td>Department of Geography, University College London, London, UK</td>
</tr>
<tr>
<td>45</td>
<td>Department of Geography and Environment, University of York, York, UK</td>
</tr>
<tr>
<td>46</td>
<td>Wildlife Conservation Society-DR Congo, Kinshasa I, Democratic Republic of Congo</td>
</tr>
<tr>
<td>47</td>
<td>Centre de Formation et de Recherche en Conservation Forestiere (CEFRECOF), Epulu, Democratic Republic of Congo</td>
</tr>
<tr>
<td>48</td>
<td>Faculté de Gestion de Ressources Naturelles Renouvelables, Université de Kisangani, Kisangani, Democratic Republic of Congo</td>
</tr>
<tr>
<td>49</td>
<td>Geography, College of Life and Environmental Sciences, University of Exeter, Exeter, UK</td>
</tr>
<tr>
<td>50</td>
<td>National Herbarium, Yaounde, Cameroon</td>
</tr>
<tr>
<td>51</td>
<td>ForestGEO, Smithsonian Tropical Research Institute, Panamá, Republic of Panama</td>
</tr>
<tr>
<td>52</td>
<td>Royal Botanic Garden Edinburgh, Edinburgh, UK</td>
</tr>
</tbody>
</table>
Service d’Évolution Biologique et écologie, Faculté des Sciences, Université Libre de Bruxelles, Brussels, Belgium

Faculty of Science, Department of Botany and Plant Physiology, University of Buea, Buea, Cameroon

Environmental Change Institute, School of Geography and the Environment, University of Oxford, Oxford, UK

Salonga National Park, Kinshasa I, DR Congo

Bureau Waardenburg, The Netherlands

Faculty of Forestry, University of British Columbia, Vancouver, Canada

Faculté des Sciences Appliquées, Université de Mbujimayi, Mbujimayi, Democratic Republic of Congo

Yale School of Forestry & Environmental Studies, New Haven, CT, USA

Department of Biological Sciences, Florida International University, FL, USA

Agence Nationale des Parcs Nationaux, Libreville, Gabon

Institut de Recherche en Ecologie Tropicale, Libreville, Gabon

School of Natural Sciences, University of Stirling, Stirling, UK
Quantifying carbon dynamics in forests is critical for understanding their role in long-term climate regulation\textsuperscript{1,2,3,4}. Yet little is known about tree longevity in tropical forests\textsuperscript{3,5,6,7,8}, a factor that is vital for estimating carbon persistence\textsuperscript{3,4}. Here we calculate mean carbon age (the period that carbon is fixed in trees\textsuperscript{7}) in different strata of African tropical forests using (i) growth-ring records with a unique timestamp accurately demarcating 66 years of growth in one site and (ii) measurements of diameter increments from the African Tropical Rainforest Observation Network (23 sites). We find that in spite of their much smaller size, in understory trees mean carbon age (74 years) is greater than in subcanopy (54 years) and canopy (57 years) trees, and similar to carbon age in emergent trees (66 years). The remarkable carbon longevity in the understory results from slow and aperiodic growth as an adaptation to limited resource availability\textsuperscript{9,10,11}. Our analysis also reveals that while the understory represents a small share (11\%) of the carbon stock\textsuperscript{12,13}, it contributes disproportionally to the forest carbon sink (20\%). We conclude that accounting for the diversity of carbon age and carbon sequestration among different forest strata is critical for effective conservation management\textsuperscript{14,15,16}, and for accurate modelling of carbon cycling\textsuperscript{4}.

Investing in carbon storage and sequestration represent important climate change mitigation strategies\textsuperscript{3}. Forests have a potential to provide both long-lived carbon stocks\textsuperscript{7,17} and long-term carbon sinks\textsuperscript{1,2}. To successfully conserve forest carbon stock and increase forest carbon uptake, we must conserve carbon-rich forests and extend the forested land area\textsuperscript{3}, but decision makers and managers also need to understand the long-term behaviour of carbon within forests\textsuperscript{1,2,3}. Critical questions are: (i) how long does the carbon persist, and (ii) where does it stay longest in the system? Carbon residence time is a direct function of tree longevity\textsuperscript{3,7,17}, but attempts to estimate tree age in tropical rainforests are relatively scarce and often highly contested\textsuperscript{5,6,7,8}. Estimating the ages of the oldest trees in tropical forest stands is particularly subject to debate. While some authors claim that broadleaved trees in the tropics may reach ages of 1000 years or older\textsuperscript{8,18}, others estimate maximum ages of not more than 600 years\textsuperscript{5,6}. Furthermore, the oldest carbon in the system is often assumed to be located in large trees\textsuperscript{8}. Canopy and emergent trees contain a large proportion of the stand-level biomass\textsuperscript{12,14} but large trees alone may not represent well the entire forest in terms of growth rates, tree lifespan and carbon persistence\textsuperscript{7}. Canopy species grow faster\textsuperscript{19}, but there is a general trade-off between
growth and lifespan in organisms\textsuperscript{9,20}. Therefore, long-term carbon storage and sequestration in tropical rainforests may substantially depend on smaller trees too.

Here, we take advantage of a remarkable rediscovery of a historic forest inventory plot to probe the age structure of African rainforests in a way that has not been possible to date. The Nkulapark plot was established in 1948 in the southwest of the Democratic Republic of the Congo (Supplementary Fig.1). A total of 6315 trees with diameter on breast height (DBH) \( \geq 5 \text{cm} \) were tagged and DBH was measured annually for 9 years. In 2014, we rediscovered and measured 450 surviving tagged trees, of which 55 were selected to measure growth ring series. We used the grown-in iron nail as a 1948 timestamp, giving accurate estimates of the DBH growth rate (in mm yr\(^{-1}\)) and the rate of growth-ring formation (number of rings per year) over a 66 year period for each tree (Fig.1, Supplementary Table 1). The age of each individual tree (in years) was calculated as the total number of rings from pith to bark, divided by the rate of growth-ring formation (number of rings per year) (Fig.1, Fig.2). We used the five-class Crown Illumination Index of Dawkins & Field (hereafter CII\textsuperscript{21,22}) to compare growth patterns and tree age among the different rainforest strata: understory (CII=1), sub-canopy (CII=2,3), canopy (CII=4) and emergent stratum (CII=5). Understory trees receive no direct light, sub-canopy trees receive lateral light, canopy and emergent trees receive vertical light (Fig.3a). Supplementary analyses show that the rediscovered Nkulapark tree dataset is adequate to compare age differences between forest strata (Supplementary Fig.2, Supplementary Fig.3, Supplementary Discussion).

We find that the age of the 55 Nkulapark trees with growth-ring series ranges between 129 and 452 years (Fig.2, Supplementary Table 1). There is no clear linear relationship between tree age and their DBH (p=0.082, Fig.2a). Understory trees (CII=1) do not differ significantly in age from canopy and emergent trees (CII=4 or 5) (p=0.254, boxplot at the right of Fig.2b), while trees in sub-canopy classes (CII=2 or 3) are slightly younger than trees in both the understory and the canopy. Despite their small size, understory trees (CII=1) can be surprisingly old. One Microdesmis puberula (TreeID=3545) has an estimated age of 329 years, with a DBH of just 156 mm (Supplementary Table 1).
To test if our findings hold true in a wider geographic context, we compared growth and age patterns among the different forest strata in 23 Central African permanent forest inventory plots¹ (Fig.3, Table 1, Supplementary Fig.1 and Supplementary Table 2). Selected plots had a similar species composition as the Nkulapark. Plots are demarcated rectangles or squares of median size 1 ha where each tree is mapped, tagged and measured according to standard protocols¹². DBH of each tree with DBH\(\geq\) 10 cm was measured at least twice. Small trees that grew larger than 10 cm during the monitoring period were recorded as recruits. Trees that died were recorded. We used repeated diameter measurements to estimate the growth rate of each individual tree. We estimated tree age by dividing the final diameter (mm) by the growth rate (mm yr\(^{-1}\)), assuming a constant growth rate over the lifetime of a tree⁷. We evaluated the robustness of this age estimation method using the rediscovered Nkulapark trees as a reference (Supplementary Fig.4). Uncertainties are due to relatively short plot monitoring periods (average 9 years), yielding negative or zero growth rates for 9.7% of the trees. To avoid unrealistic tree age estimates, we replaced slow growth rates by a ‘minimum allowed growth rate’, defined as the \(x^{th}\) percentile of the growth rate distribution within each CII class. Sensitivity analysis showed that \(x=25\) returned a realistic tree age distribution for our dataset (Supplementary Table 3, Supplementary Discussion). Further analysis suggests that \(x\) may be lower if monitoring periods are longer. Finally, we estimated tree-level mean carbon age as the average age of each year ring, weighted by the carbon content of the ring⁷, with a year ring sequence deduced from the growth rate.

The mean tree age for the 23 plots across the Congo Basin ranged between 131 and 284 years, with an overall mean of 229 years (95% bootstrapped confidence intervals: 212-244 years) (Table 1). Mean tree age in the understory (CII=1) is estimated to be 262 years, which is significantly older than the overall mean (\(p<0.001\)) and older than the mean age of the sub-canopy (CII=3, 187 years, \(p<0.001\)), the canopy (CII=4, 194 years, \(p<0.001\)) and emergent classes (CII=5, 221 years, \(p=0.021\)) (Table 1). Furthermore, mean carbon age at the tree level is 65 years (95% CI: 61-70). Carbon stored in the understory trees (CII=1) is estimated to be on average 74 years (69-79), which is significantly older than the overall mean (\(p<0.001\)) and older than carbon stored in the sub-canopy (CII=3, 54 years, \(p<0.001\)) and the canopy (CII=4, 57 years, \(p=0.001\)).
However, the difference between mean carbon age in the understory (CII=1) versus the emergent class (CII=5, 66 years) is not significant (p=0.086) (Fig.3b and Table 1).

For each forest stratum within each plot, we calculated above-ground biomass-carbon stock (AGC-stock, in Mg C ha\(^{-1}\)) and the AGC-sink (Mg C ha\(^{-1}\) yr\(^{-1}\)) using standard methods\(^1,2\) (Table 1). The AGC-stock represents the carbon reservoir in the system while the AGC-sink represents the net change, calculated as AGC-productivity (additions to the system from tree growth) minus AGC-mortality (losses)\(^{17}\). The mean AGC-stock in the understory (CII=1) is 19 Mg C ha\(^{-1}\), which is 11% of the total plot-level AGC-stock (Table 1). The mean AGC-sink in the understory (CII=1) is 0.21 Mg C ha\(^{-1}\) yr\(^{-1}\) (Fig.3c), which represents 21% of the plot-level AGC-sink. In contrast, the sub-canopy classes (CII=2 and 3) together contribute about 24% to the AGC-stock, but nothing to the AGC-sink (-0.01 Mg C ha\(^{-1}\) yr\(^{-1}\)). The understory thus contributes disproportionally to carbon sequestration, considering its relatively small share in the stock, and compared to the limited contribution of the sub-canopy classes.

**Discussion**

Results from both the Nkulapark dataset (Fig.2) and the 23 inventory plots (Fig.3, Table 1) show that understory (CII=1) and emergent (CII=5) trees are on average older than sub-canopy trees (CII=3). This pattern can be explained by differences in species composition\(^{15}\). Understory ‘specialist species’ maintain low growth rates for long periods, resulting in relatively small DBH at older age. Their adaptations allow them to survive in the understory without the need to invest in rapid growth. The sub-canopy classes (CII=2 and 3) are populated with suppressed ‘canopy specialists’ that survived a recruitment stage but didn’t reach the canopy yet. These trees are relatively young (Fig.3) and they experience limited light conditions, resulting in mortality rates equaling productivity rates (Table 1). Canopy and emergent trees (CII= 4 and 5 respectively) are canopy specialists that have been able to grow rapidly for long periods\(^{23}\). This is in line with ref.\(^{19}\) who show that growth rates in emergent trees are high and increase continuously. To test the assumption that high tree and carbon age in the understory may be due to a difference in species composition, we classified species as (i) understory specialists, (ii) non-specialists and (iii) canopy specialists
(Supplementary Discussion). This confirmed that understory specialists are on average smaller (p<0.001) but older (p=0.003) than canopy specialists (Supplementary Table 4).

Furthermore, the Nkulapark dataset shows that there is a significant negative relationship between tree age and growth-ring formation rate (Fig.2b). 91% of the trees in the dataset did not form a growth-ring every year, suggesting an aperiodic growth pattern (shifts of growth to dormancy and back do not occur annually). This aperiodic growth pattern is more prominent in understory trees, which formed fewer growth rings (0.36 rings per year) than canopy and emergent trees (0.61 rings per year) (p=0.01; boxplots at the top of Fig.2b). Aperiodic growth patterns translate into significantly slower long-term growth rates: understory trees (CII=1) have a mean DBH growth rate of 1.52 mm yr$^{-1}$ versus 4.99 mm yr$^{-1}$ in emergent trees (p<0.001) (Table 1). The observed differences in growth periodicity (Fig.2b), growth rates (Table 1) and age patterns (Fig.3) among the different forest strata are most likely a result of differences in survival mechanisms which are a function of resource availability$^{11,24,25}$ (Supplementary Discussion).

Our data show that growth and age distribution of tropical trees in mixed lowland rainforests is complex. Large canopy trees are among the oldest in the rainforest, as suggested by several authors$^8$, but as they obtained their size and position by maintaining fast growth rates$^{19}$, they are not significantly older than slow-growing small understory specialists (Fig.2, Fig.3, Table 1). Furthermore, while large canopy trees represent the largest share of the carbon stock$^{12,13,14}$, they suffer most during drought periods$^{26}$. In comparison, understory trees represent a smaller carbon capital but they are less vulnerable to drought and contribute disproportionately to carbon sequestration. As such, the understory provides long-term stability in forest carbon cycling. Furthermore, the understory is more diverse than the canopy in terms of species composition$^{14,15,16}$. Therefore, we recommend quantifying forest ecosystem services by considering forests as a whole, with all forest strata providing specific services$^{16}$. This is important in Central Africa, where the demand for fuelwood and charcoal could severely affect the understory if only large trees were to be protected$^{14,27}$. 
Finally, our results suggest that care is required with large-scale modelling of forest carbon accumulation potential and responses to different climate change scenarios. There are two hypothesised responses to increasing atmospheric CO₂ concentrations, possibly explaining the long-term observed AGC-sink in tropical forests: (i) big trees increase their asymmetric competition benefit to the detriment of the rest of the stand, or (ii) suppressed trees do best, as they are closer to their light compensation point. Our results show that both scenarios occur in forest stands, with the understory (CII=1), the canopy (CII=4) and the emergent (CII=5) classes contributing to carbon sequestration, while the sub-canopy classes (CII=2 and 3) do not contribute. Forest and carbon cycle models will need to account for the diversity of carbon age and carbon sequestration potential among the different forest strata. Recent studies have found that forest structure can be predicted from the characteristics of canopy trees, but our results suggest that temporal dynamics differ between forest strata. The long-term persistence of function depends on smaller trees too, which compared to their stature contribute disproportionately heavily to long-term carbon storage, sequestration, and climate resilience.

References


27. FAO. *The charcoal transition: greening the charcoal value chain to mitigate climate change and improve local livelihoods.* (Food and Agriculture Organization of the United Nations, 2017).


**Correspondence**

Correspondence and requests for materials should be addressed to W.H. (whubau@gmail.com).
Acknowledgments

W.H. and T.D.M. were both funded by the Brain program of the Belgian Federal Government (BR/132/A1/AFRIFORD and BR/143/A3/HERBAXYLAREDD). The PhD project of T.D.M and the tenure track of J.V.d.B. were supported by Ghent University Special Research Fund (BOF). Fieldwork was sponsored by the King Leopold III fund for nature exploration and conservation. B.A.I. is supported by the Institut National pour l’Etude et la Recherche Agronomiques en R.D.Congo (INERA- RDC- Luki) and the Ecole Régionale Postuniversitaire d'Aménagement et de Gestion intégrés des Forêts et Territoires tropicaux (ERAIFT Kinshasa). We thank WWF-RDC (Geert Lejeune), INERA and ERAIFT for facilitating fieldwork in the Luki Reserve. We thank the INERA employees (Jean-Baptiste Ndunga, Jean-Maron, Fils Mbungu Phaka, Leonard Ngoma, Noble, Plaside), the WWF ecoguards and the students of the Universities of Kinshasa (UNIKIN) and Boma for assistance in the field. For assistance with datasets we thank Marlène De Groot, Kévin Lievens, Piet Dekeyser, Stijn Willen and José Kempenaers. This paper is also a product of the AfriTRON network, for which we are indebted to hundreds of institutions, field assistants and local communities. This network has been supported by the European Research Council (291585, “T-FORCES”-Tropical Forests in the Changing Earth System, Advanced Grant to O.L.P.), the Gordon and Betty Moore Foundation, the David and Lucile Packard Foundation, the European Union’s Seventh Framework Programme (283080, ‘GEOCARBON’), and Natural Environment Research Council (NERC) Consortium Grant ‘TROBIT’ (NE/D005590/1), ‘BIO-RED’ (NE/N012542/1) and a NERC New Investigators Grant, the Royal Society, the Centre for International Forestry (CIFOR) and Gabon’s National Parks Agency (ANPN). We are indebted to the University of Yaounde I, the National Herbarium of Yaounde, Rougier-Gabon, the Marien Ngouabi University of Brazzaville, WCS-Congo, Salonga National Park, WCS-D.R.Congo, and the University of Kisangani for logistical support in Africa.

Author contributions

forest census network programmes and the Forestplots.net database, and most co-authors helped collecting AfriTRON forest census data. S.L.L., B.S., S.B., A.C.S., W.H., T.S. and L.W.W. coordinated forest plots data collection. M.J.P.S., G.L.G., S.L.L., O.L.P., T.R.B. and G.P. contributed tools to analyze and curate data. All co-authors commented on or approved the manuscript.

Competing financial interests

The authors declare no competing financial interests.

Figure legends

Figure 1 Example of a wood core (Greenwayodendron suaveolens, TreeID=765) showing the 1948 nail trace. The image at the top shows the full core. White lines indicate growth-ring boundaries, numbers indicate growth-rings (counted from bark to pith), black arrows indicate important years. The bark to the right of the figure indicates the year of sampling (2014). The dark discoloration in growth-rings 26 to 35 was caused by oxidates from the iron nail that were transported up and down in damaged vessels and fibers. The right border of the discoloration accurately marks the start of the year 1948. There are 25 rings between the bark and the 1948 nail trace, suggesting that this individual needed on average 2.6 years to form a ring. Using this rate for the 53 rings that were formed before 1948, we find that the first ring in the core was probably formed around 1811. The location of the pith is indicated by the black lines to the left, which follow the direction of the wood rays. This shows that the distance from the pith to the first ring boundary in the core is about 11 mm. When using the average ring width of rings 78 to 68, we estimate that 7 rings are missing. As such, this individual would be about 224 years old. The three close-ups at the bottom illustrate wood anatomical details used to identify growth-ring boundaries (indicated by white triangles). Ring boundaries in this species are demarcated by distended wood rays and flattening of the fibers. Black scale bars cover 0.2 mm.
Figure 2 Variation in tree age inferred from growth-ring patterns in 55 trees where nail traces accurately mark the year 1948. Panel (a) shows the relation between tree age and DBH. Panel (b) shows the relation between tree age and the growth-ring formation rate (number of rings per year) between 1948-2014. In both panels, p-values of simple linear regression models are given in red. Brown dots represent understory individuals (Crown Illumination Index =1), grey dots represent sub-canopy individuals (CII=2 and 3), blue dots represent canopy and emergent trees (CII=4 and 5). Boxplots show the first quartile, the median value and the third quartile of the tree age distribution (vertical axis, boxplots to the right) and of the variables in the x-axis (horizontal boxplots at the top of the panels). P-values under boxplots resulted from two-sided Wilcoxon rank-sum tests. Outliers are marked with open circles.

Figure 3 Estimation of mean carbon age and AGC-sink per crown illumination category, for 23 permanent inventory plots in Central Africa. The 450 rediscovered Nkulapark trees were treated as an additional plot to estimate distributions of mean carbon age. Panel (a) illustrates the Crown Illumination Index (CII) developed by Dawkins and Field (ref.22), with yellow arrows indicating reception of sunlight (drawing modified from ref.21). Understory trees (brown, CII=1) receive no direct sunlight, sub-canopy trees (grey) receive lateral (CII=2) or restricted vertical (CII=3) light, canopy trees (blue, CII=4) are almost completely exposed to vertical light and emergent trees (blue, CII=5) have a crown that is completely exposed to vertical and lateral light. Boxplots in panels (b) and (c) show the distribution of plot-level mean carbon age and plot-level AGC-sink per CII class. Boxplots represent the 25 % quartile, the median value and the 75 % quartile of the plot-level average ages. Outliers are marked with open circles. Comparison among CII classes was performed with Dunn's rank-sum test using the Benjamini-Hochberg adjustment for multiple comparisons 30. The grey dotted line shows the overall (plot-level) average age.
### Table 1

Estimation of plot-level mean tree age, mean carbon age, aboveground carbon sink (AGC-sink), AGC-stock and leaf habit per crown illumination category, for 23 permanent forest plots. The 450 rediscovered Nkulapark trees were treated as an additional plot for estimation of tree age and mean carbon age (first two rows). All metrics were averaged per plot and per CII class. Mean values are given in bold, 95% confidence intervals are given between brackets. Components of the AGC-sink are AGC-productivity and AGC-mortality. Components of AGC-stock are diameter (DBH), wood density and stem density. Comparison among CII classes was performed with Dunn's rank-sum test using the Benjamini-Hochberg adjustment for multiple comparisons; the reported p-values compare CII1 and CII5.

<table>
<thead>
<tr>
<th>metric</th>
<th>All trees</th>
<th>CII 1</th>
<th>CII 2</th>
<th>CII 3</th>
<th>CII 4</th>
<th>CII5</th>
<th>p-value</th>
</tr>
</thead>
<tbody>
<tr>
<td>Tree age (yr)</td>
<td>229 (212-244)</td>
<td>262 (243-282)</td>
<td>210 (196-223)</td>
<td>187 (170-204)</td>
<td>194 (175-212)</td>
<td>221 (192-250)</td>
<td>0.021</td>
</tr>
<tr>
<td>Mean carbon age (yr)</td>
<td>65 (61-70)</td>
<td>74 (69-79)</td>
<td>60 (55-64)</td>
<td>54 (49-59)</td>
<td>57 (51-63)</td>
<td>66 (57-75)</td>
<td>0.086</td>
</tr>
<tr>
<td>AGC-prod (MgC ha⁻¹ yr⁻¹)</td>
<td>3.1 (2.83-3.41)</td>
<td>0.47 (0.4-0.56)</td>
<td>0.53 (0.42-0.67)</td>
<td>0.35 (0.27-0.43)</td>
<td>0.9 (0.69-1.15)</td>
<td>0.82 (0.63-1)</td>
<td>0.015</td>
</tr>
<tr>
<td>AGC-mort (MgC ha⁻¹ yr⁻¹)</td>
<td>2.05 (1.66-2.48)</td>
<td>0.26 (0.19-0.36)</td>
<td>0.62 (0.49-0.76)</td>
<td>0.28 (0.18-0.38)</td>
<td>0.61 (0.41-0.85)</td>
<td>0.27 (0.07-0.61)</td>
<td>0.058</td>
</tr>
<tr>
<td>AGC-sink (MgC ha⁻¹ yr⁻¹)</td>
<td>1.05 (0.63-1.5)</td>
<td>0.21 (0.09-0.33)</td>
<td>-0.08 (-0.23-0.05)</td>
<td>0.07 (-0.06-0.2)</td>
<td>0.3 (0-0.59)</td>
<td>0.55 (0.26-0.79)</td>
<td>0.011</td>
</tr>
<tr>
<td>AGC-stock (Mg C ha⁻¹)</td>
<td>177 (159-198)</td>
<td>19 (15-23)</td>
<td>26 (21-31)</td>
<td>17 (12-22)</td>
<td>58 (42-74)</td>
<td>59 (46-75)</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>stem density (stems ha⁻¹)</td>
<td>428 (392-465)</td>
<td>194 (165-228)</td>
<td>112 (94-130)</td>
<td>38 (33-44)</td>
<td>59 (46-70)</td>
<td>25 (19-31)</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>DBH (mm)</td>
<td>313 (291-338)</td>
<td>161 (154-168)</td>
<td>235 (222-250)</td>
<td>299 (273-326)</td>
<td>435 (397-469)</td>
<td>700 (644-760)</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>DBH growth (mm yr⁻¹)</td>
<td>2.38 (2.17-2.63)</td>
<td>1.51 (1.35-1.69)</td>
<td>2.16 (1.96-2.38)</td>
<td>2.98 (2.69-3.31)</td>
<td>4.03 (3.71-4.34)</td>
<td>4.99 (4.32-5.68)</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>wood density (g cm⁻³)</td>
<td>0.59 (0.56-0.61)</td>
<td>0.64 (0.62-0.65)</td>
<td>0.61 (0.6-0.63)</td>
<td>0.59 (0.56-0.61)</td>
<td>0.59 (0.56-0.61)</td>
<td>0.57 (0.53-0.61)</td>
<td>0.006</td>
</tr>
<tr>
<td>ratio evergreen:deciduous</td>
<td>3.27 (2.85-3.71)</td>
<td>4.72 (3.86-5.59)</td>
<td>4.61 (2.76-7.85)</td>
<td>3.62 (2.78-4.6)</td>
<td>3.03 (2.09-4.22)</td>
<td>1.78 (0.88-3.43)</td>
<td>&lt;0.001</td>
</tr>
<tr>
<td>proportion evergreen trees (%)</td>
<td>55 (52-57)</td>
<td>56 (52-59)</td>
<td>55 (51-59)</td>
<td>56 (52-61)</td>
<td>51 (46-56)</td>
<td>44 (37-52)</td>
<td>0.01</td>
</tr>
<tr>
<td>proportion deciduous trees (%)</td>
<td>18 (16-20)</td>
<td>14 (12-17)</td>
<td>19 (16-22)</td>
<td>21 (17-25)</td>
<td>26 (21-31)</td>
<td>35 (28-41)</td>
<td>&lt;0.001</td>
</tr>
</tbody>
</table>
Online Methods

Site description. The Nkulapark is a phenology and tree-growth monitoring plot covering 174 ha within the Luki Man and Biosphere reserve, located in the southern Mayumbe mountains in the Democratic Republic of the Congo (Supplementary Fig.1). The region experiences a humid tropical climate with a dry season between mid-May and mid-October and a short dry season from mid-December to mid-February. Yearly precipitation ranges from 649 mm to 1853 mm with a mean precipitation of 1173 mm. Temperature ranges between 19 °C and 30 °C with a mean temperature of 25.5 °C. The Nkulapark is situated almost entirely in a catchment with a valley and a ridge and includes several microclimatic conditions. The semi-deciduous lowland forest consists of (i) mature forest dominated by Prioria balsamifera, (ii) old regenerating forest dominated by Terminalia superba, (iii) mixed-species mature forest and (iv) modified forest patches.

Nkulapark plot design in 1948. The Nkulapark was established and managed by the Institut National pour l’Etude Agronomique du Congo Belge (INEAC), which was later renamed Institut National pour l’Etude et la Recherche Agronomique en R.D.Congo (INERA, http://www.inera-drc.org). The person in charge of the tagging and the measurements was Léon Toussaint, who worked as a botanist in the Luki reserve between 1946 and 1952. The planning of the plot was first announced in the INEAC-Luki annual report of 1946. A total of 29.2 km of observation paths were cut in the forest in 1947, following the contour lines of the Nkula river valley (Supplementary Fig.1). A total number of 6315 trees were tagged by the end of 1947, so we assume that first wood formation after tagging occurred during the wet season that started in October 1948. Tree selection was performed by randomly selecting trees from the pool of trees > 5 cm DBH in 1948. The Nkulapark area was mapped, showing the locations of the largest tagged trees. From 1948 to 1957 yearly diameter measurements were performed on all tagged trees. Mortality events were recorded in the datasheets. Trees were measured at a height of 1.3 m and the point of measurement (POM) was indicated on the tree with a horizontal line of lead-based paint. For trees with buttresses or deformities, the POM was raised to a point high enough to avoid the irregularities interfering with diameter measurements at subsequent censuses. For trees with extremely high buttresses, diameters were estimated. For the same 6315 trees, weekly phenology observations were recorded. Phenological observations were done for leaf habit,
flowering, fruiting and seed dissemination. The plot was abandoned in 1957 but the datasheets were kept in the library of the INERA station in Luki and digitized in 2008 and 2014.

Rediscovering Nkulapark trees in 2014. Each of the original 6315 Nkulapark trees were labelled during the first census in 1948 with a zinc number tag that was attached to the tree using an iron nail of 8 cm long. A part of these trees were indicated on the original 1948 map. During a first prospective field campaign in August 2014, this map was digitized and georeferenced with QGIS (QGIS development team, 2016) using landmarks such as easily rediscovered trees, contour lines and observation paths that were still visible and could be tracked with a Garmin 64s GPSmap (see Supplementary Fig.1). Based on this map, we pinpointed the approximate location of 1521 individuals that were recorded as alive during the last census in 1957. During a second field campaign in September-October 2014, these 1521 individuals were searched for. 450 of them were found alive, 16 were found dead and the remaining 1057 could not be relocated and were assumed to be dead and rotten away, albeit some may have been missed (see Supplementary Fig.2, Supplementary Fig.3, Supplementary Discussion for an in-depth analysis of survivorship rates). The original 1948 tree tags and nails of the rediscovered trees were either still present outside the trunk or detected inside the tree using a metal detector (BHJS, Bounty Hunter, USA). Scars on the trunk indicated the presence of an overgrown nail, and repelled number tags were sometimes found on or in the ground nearby the tree using the metal detector. In most cases, the numbers on the tags were still readable. On 95% of the rediscovered trees, the lead-based paint of the POM was still visible, allowing a representative DBH measurement. Comparison of DBH, DBH growth rates and tree age distribution in the original dataset of Nkulapark trees versus the dataset of rediscovered trees, showed that the rediscovered dataset is slightly biased towards discovering slower-growing trees, but this bias affected both the classes of understory and canopy ‘specialist’ species (Supplementary Fig.3). Hence, the rediscovered tree dataset is representative to compare growth and age patterns among the different forest strata in the Nkulapark area.

Sampling Nkulapark trees in 2014. Wood samples for growth-ring analysis were taken from rediscovered trees if following criteria were met: (a) the nail was still present in the wood, either totally grown-in or partly sticking out of the trunk, (b) the exact position of the nail could be identified visually or with the metal
detector, (c) the nail was not overgrown by excessive wound tissue, buttresses or other deformities. As such, increment cores or stem discs were taken near the nail for 58 of the rediscovered trees. For each sampled tree, increment cores were taken a few centimetres above, below, to the left and to the right of the nail using a 40cm Pressler bore. For each tree, two additional cores were taken at 120° from the nail trace along the circumference of the tree. As such, 6 increment cores were available for each tree. This maximised the chance of sampling the pith of the tree. To study and describe the reaction of the wood after tagging, additional larger wood samples containing the nail were extracted from 30 trees using a saw and a chisel.

Visualizing and measuring growth-ring series. For each wood core, growth-ring series were visualized using two imaging methods as described by ref. 36: (i) first, density profiles were calculated from X-ray CT scans of entire wood cores, then (ii) the cores were surfaced with a core microtome and scanned using a flatbed scanner (EPSON Perfection 4990 PHOTO). To obtain X-ray CT volumes, cores were scanned at 110 µm resolution with the Nanowood CT facility from the Centre for X-ray Computed Tomography of Ghent University (UGCT, www.ugct.ugent.be), developed in collaboration with XRE (www.xre.be; now part of the TESCAN ORSAY HOLDING a.s.). Reconstruction was performed with the Octopus software package on a GeForce GPX 770 4GB GPU. X-ray and flatbed scans were analysed using the toolchain for tree-ring analysis described by De Mil et al. (ref. 36). This toolchain semi-automatically indicates the growth-ring boundaries and calculates growth-ring width series. Depending on the visibility of the growth-ring patterns, either the X-ray or the flatbed scans were used to check growth-ring boundaries and measure growth-ring widths. Growth-ring boundaries were distinguished using visual wood anatomical characteristics such as distended rays, flattened fibers and terminal parenchyma bands. For unclear ring boundaries, microscopic thin sections were taken to study wood anatomy at high resolution using an Olympus BX60 microscope (Fig.1).

Detecting the 1948 nail trace. Discolorations or wound tissue formed as a reaction on the tagging were visible in the cores taken near the nail. The surface of these stem discs was sanded to a few millimetres above the nail and the anatomy was observed using an Olympus BX60 microscope. The nails were remarkably well preserved inside the trees, probably due to cathodic protection of the iron by the zinc of the
tags. Hence, discolorations or wound tissue were visible in each of these samples. Discolorations were recognizable as darkened tissue in otherwise light-coloured wood (Fig.1). Discolorations occurred in cells that were formed before the nail was inserted, due to oxidation processes between the wood and the iron nail. Water in vessels and fibers in the neighbourhood of the nail (especially those damaged by the nail) spread these oxidates upwards and downwards. Therefore, the discolorations in these cells are also detectable on samples that were taken a few centimetres under or above the nails. Vessels and fibers that were formed after the tagging were not damaged, hence did not show discolorations. As such, the boundary of the discoloration accurately serves as a timestamp indicating the year of tagging (1948). Furthermore, wound-induced deformities occurred in the wood that was formed after the nail was inserted. This wound tissue is characterized by increased woody productivity around the nail, forming a lump in the growth-rings that were formed just after the nail was inserted. This lump formation is not present in the wood that was formed before tagging.

Estimating tree age using growth rings and nail traces. Six cores were assessed for each tree. For some trees, none of the cores contained the pith because (i) the tree radius exceeded the borer length or (ii) the pith was eccentric and missed. In these cases, the core with the largest number of visible rings was used to estimate the total number of growth-rings. The missing core length from the end of this core to the pith was estimated using the intersection of three lines of ring boundaries marked along the rays, as described by ref.29 (see Fig.1 for an example). We estimated the number of rings in the missing core part by dividing the missing core length by the average ring width of the 5 oldest rings of the sampled core. We tested the robustness of this method by rerunning the analysis using the average growth rates of the 5, 10, 15 and 20 oldest rings. We found that the overall average tree age and the trends observed in Fig.2 are not sensitive to varying the method used to estimate the number of rings in the missing part of the core. The core with the clearest nail trace was used to count the number of rings formed after the year of tagging (1948). We used the number of growth-rings formed between 1948 (nail mark) and 2014 (cambium) as a reference to calculate the number of years the individual needed to form one ring (years per ring). We then multiplied this ratio with the total number of rings formed before 1948 to estimate the age of the tree (Fig.1 and Fig.2). Three individuals were
not used for analysis because the estimated missing core length of each core exceeded 20 cm. As such, growth-ring series of 55 trees were retained (Supplementary Table 1).

Permanent forest inventory plots. To test if our findings hold true in a wider geographic context, we estimated mean tree age and carbon age of the different forest strata in 23 Central African permanent forest plots located in four different Central African countries (Cameroon, Gabon, Congo Brazzaville, D.R.Congo) (Supplementary Fig.1 and Supplementary Table 2). Plots were selected if at least 65% of the trees belonged to species that also occur in the Nkulapark. Furthermore, plots selected for analysis conformed to the following criteria: (1) plots had an actual plot area of ≥0.2 ha, (2) plots were georeferenced, (3) all trees with DBH ≥100 mm were measured, (4) the majority of stems were identified to species level, (5) plots had at least 2 censuses, (6) plots had a total monitoring length of ≥3 years, (7) plots were situated within structurally intact, apparently mature forest (excluding young or open forests), (8) plots were free from major human impacts, (9) plots were located at ≥50m from the anthropogenic forest edge, (10) altitude was below 1500 m.a.s.l., (11) mean annual air temperature was ≥20.0°C, (12) mean annual precipitation was ≥1000 mm yr⁻¹, (13) plots were located in terra firme forest. For analysis purposes, plots smaller than 0.5 ha that were within 1 km of each other and located in similar forest types were merged (i.e. the LME cluster). All selected plots are part of the African Tropical Rainforest Observatory Network (AfriTRON; www.afritron.org). These data are curated in the ForestPlots.net database, and subject to identical quality control and quality assurance procedures. All calculations of plot data metrics described hereafter were performed using the R statistical platform.

Estimating AGC-stock. For each tree and each census, Aboveground Biomass at the tree level (AGB, Mg stem⁻¹) was estimated using a published allometric equation for moist forests including terms for diameter (DBH, mm), dry wood density (ρ, g cm⁻³) and total tree height (H, m):

\[ AGB = \frac{0.0673 \times \rho \times (\frac{DBH^2}{100}) \times H^{0.976}}{1000} \] (formula 1).

Wood density values were derived from the dryad database (www.datadryad.org). Stems were matched to species-specific wood density values or the mean values for the genus or family, following ref.¹ and ⁴³.
Heights were calculated using a single height-diameter model (Weibull) for central African lowland terra firme forests published by ref. 46, using commands implemented in the R-package BiomasaFP 47. Aboveground Biomass-Carbon (referred to as AGC) is considered as 47% of the AGB following IPCC recommendations 48. For each individual tree in the plot dataset, AGC-stock was calculated as the mean of the first and last censuses.

**Estimating AGC-sink.** For the calculation of AGC-sink, only the first and the last censuses were used for each plot. First, AGC-productivity (Mg C stem\(^{-1}\) yr\(^{-1}\)) for each stem surviving the monitoring period, was calculated as the difference between its total AGC at the end census minus the total AGC at the start census of the interval, divided by the census interval length (yr). AGC-productivity for stems recruited during the monitoring period (i.e. reaching DBH ≥ 100 mm), was calculated in the same way, assuming DBH = 0 mm at the start of the interval. AGC-mortality for each tree that died during the monitoring period (Mg C tree\(^{-1}\) yr\(^{-1}\)) was calculated as the AGC at the start of the monitoring period, divided by the total monitoring length (yr). AGC-productivity at the stand level (Mg C ha\(^{-1}\) yr\(^{-1}\)) was then calculated as the sum of tree-level productivity estimates of all survivors and recruits. AGC-mortality at the stand level (Mg C ha\(^{-1}\) yr\(^{-1}\)) was calculated as the sum of tree-level mortality estimates of all dead trees in the subset. We corrected for unobserved components of biomass growth and mortality due to census interval length effects, as discussed by ref. 2, ref. 49 and ref. 50. A method to correct productivity and mortality rates for these uncertainties was developed by ref. 49 and applied by ref. 2. This correction accounts for (i) trees that recruit and die within the same interval (i.e. unobserved recruits) and (ii) growth of trees that grow and die within the interval (i.e. unobserved biomass growth from mortality, which is not recorded because dead trees are not measured). As such, for each census interval, we calculated unobserved recruitment and unobserved mortality components (Mg ha\(^{-1}\) yr\(^{-1}\)) using the formulas proposed by ref. 49 and both components were added to both the AGC-productivity and AGC-mortality estimates. Estimates of the unobserved biomass components usually accounted for less than 3% of the total AGC-productivity and AGC-mortality. Finally, the AGC-sink was calculated as stand-level AGC-productivity minus AGC-mortality. Commands to calculate AGC-stock and AGC-sink are implemented in the BiomasaFP R package 47.
Tree age inferred from DBH growth rates in permanent inventory plots. For each tree within the permanent forest inventory plots, we estimated the age by dividing the DBH (mm) in the final census with the DBH growth rate (mm yr\(^{-1}\)) of the tree itself. For each tree, we averaged the DBH growth rate over all census intervals preceding the last census. This method uses the actual growth rate of each tree, which is accurate for healthy trees but returns unrealistic age estimates for trees with (i) slightly negative growth rates, (ii) zero growth rates or (iii) very slow growth rates. Such slow growth rates may be recorded in all DBH classes. First, slow growth rates in large DBH classes may occur when a tree is diseased or at the end of its life (senescence). These growth rates may not be representative for the total lifespan of these trees. Secondly, small growth rates may be recorded for small suppressed trees if their growth is so slow that it cannot be recorded with sufficient precision using standard census procedures. Hence, these growth rates are replaced by growth rates that are comparably small but yield a realistic tree age estimate. As such, we chose a ‘minimum allowed growth rate’ per CII class, following ref.\(^7\). We calculated the minimum allowed growth rate for each CII class as the 25\(^{th}\) percentile (=first quartile) of the growth rate distribution within the CII class. For each tree with a growth rate slower than the minimum allowed rate, we replaced the growth rate by the minimum allowed growth rate of the CII class. We conducted a sensitivity analysis to check how results vary when varying the minimum allowed growth rate: we reran the analysis using the 10\(^{th}\), 15\(^{th}\), 20\(^{th}\), 25\(^{th}\) and 30\(^{th}\) percentile of the growth rate distribution within each CII class as a minimum allowed growth rate (Supplementary Discussion and Supplementary Table 3). We used the average tree age in the dataset of 450 rediscovered Nkulapark trees as a reference to evaluate the tree-age estimation method based on DBH growth rates (Supplementary Figure 4).

Mean carbon age. As a tree grows, it increasingly stores more carbon. Carbon near the bark of the tree is younger than carbon in the pith. As such, mean carbon age of a tree does not equal total tree age. We calculated mean tree-level carbon age using the same formula as ref.\(^7\):

\[
\text{mean carbon age} = \frac{\sum_{i=1}^{n} (C_i \times A_i)}{\sum_{i=1}^{n} (A_i)} \quad \text{(formula 2)}
\]

with:

\(C_i\) = carbon content of the \(i^{th}\) ring (kg).
The carbon content of the $i^{th}$ ring is calculated as the carbon content of a tree with the DBH of ring $i$ minus the carbon content of a tree with the DBH of ring $i-1$. To estimate carbon age in trees without measured growth-ring series (permanent inventory plots), we used the final DBH and the DBH growth rate (in mm yr$^{-1}$) to deduce a year-ring pattern, assuming that the growth rate was constant over the lifetime of the tree.

Classification of tree species and statistical analysis. We used the Nkulapark phenology data published by ref.24 to classify tree species as evergreen or deciduous. To distinguish between understory, sub-canopy, canopy and emergent trees in the Nkulapark and the permanent forest plots, we used the Crown Illumination Index (CII) of Dawkins & Field (ref.22). Fig.3 illustrates the 5 classes with a drawing modified from ref.21. The Crown Illumination Index was recorded in 23 of the selected permanent inventory plots. In each plot where the index was recorded, each tree was attributed to one of the CII classes. The index was attributed in the field, mostly during one census and mostly by a single person. We estimated mean tree age, mean carbon age, mean AGC-stock and mean AGC-sink for each of the CII classes in each plot where the index was recorded. None of the metrics reported in Table 1 meet the criterion of homogeneity of variances (Bartlett test). Therefore, differences among the CII classes were tested using the non-parametric Dunn's rank-sum test. To avoid the multiple comparison problem, we used the Benjamini-Hochberg p-value adjustment (‘dunn.test’ package in R).

Data availability. The input data and R-scripts to generate the figures and tables are available for download using the following private link: https://figshare.com/s/06c793575d3b52ef5574. Images of wood cores are available using the following link: https://figshare.com/s/e6101fe7d330f8ea140a. This folder also contains all annotation documents needed to visualize growth ring boundaries on the wood samples (please consult the README document for guidelines). Wood samples used to conduct this analysis are stored in the Tervuren xylarium (http://www.africamuseum.be/collections/browsecollections/naturalsciences/earth/xylarium). These samples
may be studied, within the Tervuren xylarium, upon request addressed to the curator H.B. 
(hans.beeckman@africamuseum.be) or the corresponding author W.H. (whubau@gmail.com).

References (methods only)


34. INEAC. *Rapport Annuel INEAC-Luki.* (1946).


45. Chave, J. et al. Improved allometric models to estimate the aboveground biomass of tropical trees. 


   downloaded from ForestPlots.net. R package version 1.1. Available at 


49. Talbot, J. et al. Methods to estimate aboveground wood productivity from long-term forest inventory 

a  

Tree Age (yr) from growth-rings

slp=0.08  p=0.082

0  100  200  300  400  500  600  700  800  900  1000

DBH (mm)

b  

Tree Age (yr) from growth-rings

slp=−128.02  p<0.001

0.0  0.2  0.4  0.6  0.8  1.0

Growth-ring formation rate (rings yr⁻¹)